



Distribution and fate of organochlorine pesticide residues in sediments from the selected rivers in Taiwan

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Abstract

The contamination of organochlorine pesticides (OCPs) in sediments from selected rivers in Taiwan was investigated to evaluate the pollution potentials and hazard in river sediments. Da-han River and Erh-jen River were selected as the target rivers due to their serious pollution. A total of 40 surface sediment samples were collected at five sampling stations along the rivers. Results showed that the concentrations of various pesticides in sediments were in the range of 0.57–14.1 ng/g for \sum HCH, 0.05–0.15 ng/g for aldrin, 0.12–5.8 ng/g for dieldrin, 0.22–0.64 for endrin, 0.24–6.37 ng/g for endosulfan and 0.21–8.81 ng/g for \sum DDT (*p,p'*-DDD, *p,p'*-DDE, *p,p'*-DDT). Among the OCPs, \sum HCH, endosulfan and \sum DDT were the most dominant compounds in the river sediments. Endosulfan sulfate was the most frequent detected compound in the sediments from the selected rivers. Also, \sum DDT, dieldrin and β -HCH were in abundance. Different contamination patterns between the selected river sediments were also observed. Da-han River was mainly contaminated with endosulfan sulfate and \sum DDT. Whereas the main pesticides in Erh-jen River were β -HCH and \sum DDT. Among the cyclodiene compounds, dieldrin was in abundance in most of the sediments. Moreover, the frequencies of detection of the metabolites were higher than those of parent compounds, depicting that the sediments have contaminated for a long time. The results obtained in this study showed that there still exist a variety of OCP residues in the river sediments in Taiwan. © 2002 Elsevier Science Ltd. All rights reserved.

Keywords: Organochlorine pesticides; Surface sediment; HCH; DDT; Distribution

1. Introduction

Organochlorine pesticides (OCPs) are known for their environmental persistence and global concerns. Residues of OCPs continue to detect in many areas. Owing to their high capacities for bioaccumulation and toxicities in organisms, these compounds pose threats to ecosystems and human health. Since these compounds have an

affinity for particulate matters, one of their major sinks is thought to be the river and marine sediments. Therefore, the investigation of distribution of OCPs in sediments can provide a valuable record of contamination in the aquatic environments.

The input pathways of OCPs into the river environments include discharge of domestic sewage and industrial wastewater, runoff from nonpoint sources, and direct dumping of wastes into the river. Although sediments do not constitute a direct measure of the degree of aquatic pollution, they offer an irreplaceable aid in reconstructing the historical inputs of OCPs based on profile descriptions of contamination in core samples

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(Hendy and Peake, 1996). The distribution of various contaminants in sediments depends largely on the physico-chemical properties of the ecosystem, the partition coefficients of individual contaminants, the organic contents, and microflora activities (Pavoni et al., 1987; Brook and Moore, 1988; McKenzie-Smith et al., 1994; Glynn et al., 1995). Although the residue levels of the chlorinated compounds in the environments have considerably declined in the past 20 years, recent work has depicted that chlorinated pesticides could be detected in the range of 0.03–25.17 ng/g dry weight (d.w.) in marine and river sediments (Gold-Bouchot et al., 1995; Sarkar et al., 1997). Moreover, some OCPs, such as DDT, HCH and endosulfan are still used in some countries around the tropical and subtropical belts for agricultural and medicinal purposes. These compounds can be deposited into the sediments through long-range atmospheric transport (LRT), resulting in a high exposure to OCPs in the areas near the pollution sources.

Taiwan has a strong agricultural sector, which has made wide use of pesticides. The OCPs were employed extensively in the early 1950s for the control of soil-dwelling insects and have been officially banned since 1974. It is estimated that 2.5×10^7 kg of OCPs was released into the environments annually from the 1950s to 1970s and constituted a long-term source of contamination to aquatic ecosystem via soil erosion and agricultural runoff. A variety of OCP residues in foods from Hsinchu, Taiwan were also detected in the range of 0.26–10.22 ng/g wet weight (Doong and Lee, 1999). However, there is little information of the contamination of sediments in Taiwan rivers.

The purpose of this study was to determine the contamination level of OCPs in sediments from selected rivers to evaluate the pollution potentials and hazard of river sediments in Taiwan. Two different rivers, Da-han River and Erh-jen River were selected as the target rivers for studies in northern and southern Taiwan, respectively.

2. Methods and materials

2.1. Study area and sampling

The Da-han River, one of the important rivers in northern Taiwan, has a catchment basin of 1163 km² and a total length of 135 km. The Da-han River is a traditional agricultural production site which has small scale farms producing grains. Also, this river is a main source for water supply in northern Taiwan and receives massive discharge of domestic sewage from the urbanized areas in its catchment basin. The Erh-jen River, a representative polluted river in southern Taiwan, has a catchment area of 350 km² and a total length of 65 km.

The river water is highly polluted with heavy metals and dioxin from the industrial discharge.

A total of 40 sediment samples were collected from these two rivers from October 1997 to April 1998. Five sampling stations along each river were selected and four samples from each station were collected. The locations of these sampling stations in the selected rivers are shown in Fig. 1. The distances of sampling sites from the outlets of the selected rivers were 19.3, 22.9, 23.6, 27.2 and 41.2 km for Da-han River and 1.2, 6.3, 17.9, 24.6, and 30.1 km for Erh-jen River. Water depths of the stations ranged from 0.4 to 6.7 m. The upper 15 cm of the surface sediments were sampled from the sampling sites using a boat with a Birge–Ekman sediment grabber in areas of low flow velocity (<0.3 m/s). Immediately after collection, sediments were preserved in deep freeze in order to avoid degradation.

2.2. Extraction and cleanup

Homogenized subsamples were freeze-dried and OCPs were extracted with Soxhlet apparatus. A 10 g sediment was placed into a thimble filter and OCP residues in sediments were extracted with 250 ml hexane for 24 h at a rate of 10–12 cycles/h. The extract was then preconcentrated to 2–3 ml on a rotary evaporator. The extract was further treated with activated copper powder to remove sulfur compounds. A 2 g Florisil SPE cartridge was used to cleanup the extracts. Sodium sulfate (ca. 1.0 cm) was added to a Florisil SPE cartridge. The cartridge was washed with 6 ml petroleum ether–ethyl ether (95 + 5) at a rate of 5 ml/min and then the organochlorine pesticides were eluted with 12 ml petroleum ether–ethyl ether (95 + 5) at a rate of 2 ml/min. The elutes were concentrated to about 1–2 ml on a rotary evaporator and then transferred to 10 ml glass tubes with small amounts of hexane. The solvent in the glass tube was entirely evaporated under a gentle stream of nitrogen and the precipitates were redissolved in 1 ml hexane and analyzed with a gas chromatography (GC)–electron capture detector (ECD) system.

2.3. Analytical procedures

The concentrations of OCPs in the extracts were monitored with GC equipped with an ECD and a PTE-5 fused silica capillary column (30 m length \times 0.32 mm inner diameter \times 0.25 μ m film thickness, Supelco Inc., Co., Bellefonte, PA.). One μ l of each sample was injected into the GC system for separating the OCPs. Column temperature increased from 140 to 200 °C at 15 °C/min, held for 2 min, and then programmed to 250 °C at 2 °C/min, held for 2 min. The temperatures of injector and detector were 250 and 300 °C, respectively. Nitrogen gas was used as the carrier and make-up gases at flow rate of

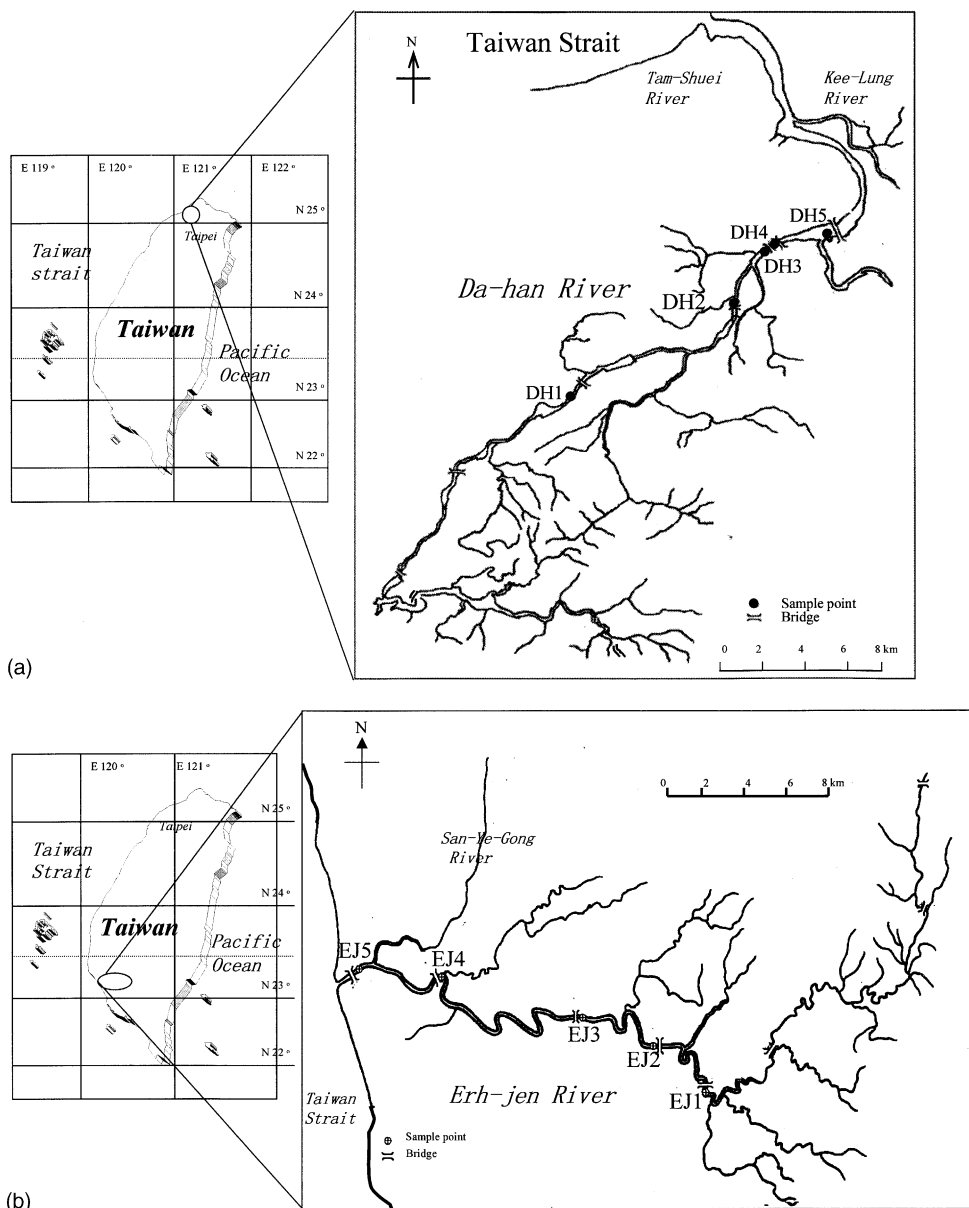


Fig. 1. The study areas and sampling locations in the selected rivers in Taiwan.

1.73 ml/min (linear velocity, 25.2 cm/s) and 35.5 ml/min, respectively. Pentachloronitrobenzene and decachlorobiphenyl were used as the internal standards. Also, identified peaks were checked by GC-ECD with SPB-608 fused silica capillary column (30 m length \times 0.53 mm inner diameter \times 0.5 μ m film thickness, Supelco Inc., Co., Bellefonte, PA). The residues of OCPs were determined by comparing the peak areas of the samples and the calibration curves of the standards. The correlation coefficients (R) of calibration curves of OCPs were all greater than 0.998.

Oxidation–reduction potential (ORP), volatile suspended solid (VSS), total organic carbon (TOC), water content and particle size distribution of sediments were also analyzed. ORP was determined with an Orion EA 920 expandable ion analyzer (Orion Research Inc., Boston., MA.) by using an Orion model 97-78-00 Pt redox electrode. The values were expressed in terms of the ORPs of the samples relative to the Pt redox electrode and were read out until the potential changed was less than 0.5 mV/min. Water content and VSS were determined by the weight loss at 103 ± 2 °C and 550 ± 50 °C,

Table 1
The recoveries, MDLs, and the analyzed results of CRM standards (840-050) of OCPs in sediments

Pesticides	Recovery (%)	MDLs (ng/g dry weight)	CRM standard (mg/kg)	
			Certified value	Analyzed value
α -HCH	85.0	0.12	– ^a	–
β -HCH	94.4	0.35	–	–
γ -HCH	83.1	0.25	491.4	470.6
δ -HCH	99.9	0.11	–	–
Heptachlor	92.0	0.15	–	–
Aldrin	92.2	0.05	18.04	19.18
Heptachlor epoxide	96.9	0.15	–	–
Endosulfan I	90.1	0.08	1464.3	1232.6
<i>p,p'</i> -DDE	88.6	0.12	1519.6	1269.6
Dieldrin	98.9	0.12	1862.5	1693.9
Endrin	121.1	0.22	62.15	67.7
Endosulfan II	96.2	0.16	1128.2	1284.3
<i>p,p'</i> -DDD	117.7	0.18	1530.6	1989.8
Endosulfan sulfate	85.2	0.13	–	–
<i>p,p'</i> -DDT	116.4	0.18	1060.1	1067.7
Endrin ketone	69.5	0.13	–	–
Methoxychlor	129.4	0.30	–	–

^a Not available.

respectively. Moreover, ampule TOC concentrations of sediment particles were obtained by a model 700 TOC analyzer using wet oxidation method (O.I. Corporation, Texas).

2.4. Quality assurance

For every set of 10 samples, a procedural blank and spike sample consisting of all reagents was run to check for interference and cross contamination. The method detection limits (MDLs) of OCPs were determined as the concentrations of analytes in a sample that gives rise to a peak with a signal-to-noise ratio (S/N) of 3. The OCP recoveries were determined relative to the ratio of direct injection of extract and the working standards prepared in hexane. Also, the quality of the analytical data was assured using the CRM 804-050 soil standard. Table 1 illustrates the recoveries and MDLs of OCPs. The recoveries and MDLs of OCPs ranged from 69.5% to 129.4% and 0.05–0.35 ng/g dry sediment, respectively. Moreover, the average analyzed concentrations of OCPs in CRM sample were in the range of 19.2–1989.8 mg/kg, which corresponding to the recoveries of 84–130%.

3. Results and discussion

3.1. Characteristics of sediments

Table 2 illustrates the basic physico-chemical parameters of sediments collected from Da-han and Erh-jen Rivers. These parameters included ORP, VSS, TOC, water content and particle size distribution. The char-

acteristics of sediments were site-dependent. The ORP values ranged from 184 to –372 mV and from –59 to –269 mV for Da-han and Erh-jen Rivers, respectively. These measured ORP values reflect that the collected sediments were mainly under anaerobic conditions (Parsons and Barrio-Lage, 1985). The distributions of particle sizes in sediments from Da-han and Erh-jen Rivers were also different. The particle sizes in Da-han River ranged mainly from 180 to 300 μm , while small-size particles (8–37 μm) were dominant in Erh-jen River. Also, the TOC contents of sediments ranged from 0.51 to 7.86 mgC/g and from 1.78 to 45.21 mgC/g for Da-han River, and Erh-jen River, respectively. It is clear that Erh-jen River was composed of fine particles and had high organic contents in the sediments. This observation is consistent with other studies, which demonstrated that fine particles can retain large amounts of organic compounds and pose a high pollution potency (Sarkar et al., 1997). Moreover, the VSS contents of the sediments ranged from 1.07% to 2.5% and from 0.71% to 3.0% for Da-han and Erh-jen Rivers, respectively. A positive correlation ($r^2 = 0.692$) between VSS and TOC in sediments was established, depicting that the sediment was contaminated with organic compounds (Fig. 2).

3.2. Concentration profiles of OCPs

Tables 3 and 4 illustrate the concentrations of OCP residues in surface sediments from Da-han and Erh-jen Rivers, respectively. Concentrations of OCPs in sediments from Da-han River were <0.12–4.94 ng/g d.w. for $\sum\text{HCH}$, <0.05–0.15 ng/g d.w. for aldrin, <0.12–5.8 ng/g d.w. for dieldrin, <0.08–3.78 ng/g d.w. for endosulfan,

Table 2
The physicochemical properties of surface sediments collected from the selected rivers in Taiwan

Sampling stations*	Water depth (m)	ORP value (mV)	VSS (%)	Water content (%)	TOC (mg C/g)	Particle size distribution (μm)
DH-1	3.2–6.7	(184)–(–184)	1.56–2.50	33.2–47.1	4.72–7.88	8–37
DH-2	1.9–4.0	(–181)–(–220)	1.52–1.82	26.2–31.9	1.76–5.32	37–300
DH-3	0.9–2.2	(–18)–(–112)	1.35–1.57	24.9–27.0	1.62–3.23	180–300
DH-4	2.0–4.5	(–172)–(–372)	1.36–2.44	23.9–51.9	1.49–5.98	8–300
DH-5	1.9–6.5	(–225)–(–290)	1.07–1.59	26.2–30.3	0.51–5.21	105–300
EJ-1	0.4–1.0	(–59)–(–185)	1.29–1.94	26.2–35.7	2.82–6.94	8–37
EJ-2	0.6–1.8	(–105)–(–234)	1.34–1.79	23.1–47.1	1.78–5.78	8–37
EJ-3	0.5–3.0	(–234)–(–255)	0.49–1.89	25.4–39.6	1.79–3.59	8–180
EJ-4	1.8–2.1	(–255)–(–269)	0.71–2.71	47.9–50.3	2.2–11.15	4–180
EJ-5	2.5–3.7	(–179)–(–269)	0.82–3.0	30.3–72.7	9.03–45.21	2–37

* Five sampling stations along each river were selected and four samples from each station were collected.

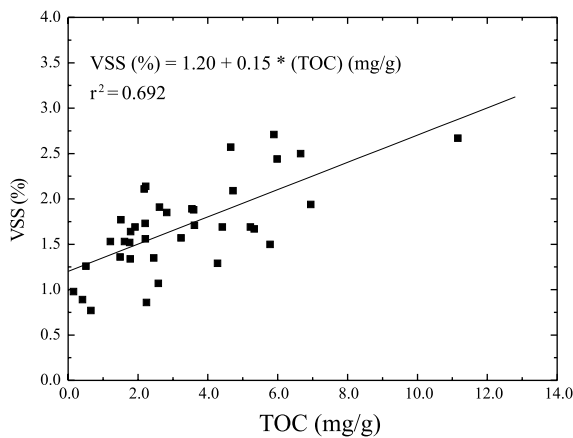


Fig. 2. The relationship between ampule TOC (mg C/g) and VSS (%) in sediments collected from the selected rivers.

<0.18–2.64 ng/g d.w. for *p,p'*-DDT and <0.3–2.85 ng/g d.w. for methoxychlor. Several biological metabolites of the parent OCPs were also detected. The concentrations were 0.32–2.39 ng/g d.w. for endrin aldehyde, 0.62–3.18 ng/g d.w. for endosulfan sulfate, 0.51–3.89 ng/g d.w. for *p,p'*-DDE and 0.56–3.34 ng/g d.w. for *p,p'*-DDD.

Concentrations of OCP residues in sediments from Erh-jen River were higher than those from Da-han River. As illustrated in Table 4, concentrations of OCPs ranged from <0.35 to 14.1 ng/g d.w. for β -HCH, <0.12 to 1.29 ng/g d.w. for dieldrin, <0.08 to 6.27 ng/g d.w. for endosulfan, <0.12 to 1.69 ng/g d.w. for *p,p'*-DDE, <0.18 to 3.9 ng/g d.w. for *p,p'*-DDD, <0.18 to 5.57 ng/g d.w. for *p,p'*-DDT and <0.3 to 7.39 ng/g d.w. for methoxychlor. This may be due to the high organic contents of the sediment particles in Erh-jen River. Several studies have depicted that the contamination of hydrophobic organic compounds in sediments is dependent on the chemical properties of the ecosystem, the partition coefficients of individual compounds and the organic

contents of sediment particles (McKenzie-Smith et al., 1994; Glynn et al., 1995). A similar distribution pattern of OCPs in sediments from the selected rivers was obtained when the concentrations of OCP residues were normalized to TOC contents. Table 5 illustrates the concentration ranges, medians and mean values of OCPs in the selected rivers based on TOC concentrations. The total OCP residues in sediments from Da-han River ranged from 0.21 to 7.42 $\mu\text{g/g}$ TOC with the mean value of 2.34 $\mu\text{g/g}$ TOC. Also, $\sum\text{HCH}$, $\sum\text{cyclodiene}$ and $\sum\text{DDT}$ concentrations ranged from ND to 2.08, from ND to 4.51 and from 0.06 to 2.91 $\mu\text{g/g}$ TOC, respectively. The normalized OCP concentrations in sediments collected from Erh-jen River was only slightly higher than that from Da-han River. The measured $\sum\text{HCH}$ ranged from ND to 3.47 $\mu\text{g/g}$ TOC with the mean value of 1.03 $\mu\text{g/g}$ TOC. Although this value is 2-fold higher than that in Da-han River, the concentration ranges and mean values of $\sum\text{DDT}$ and $\sum\text{cyclodiene}$ obtained from Erh-jen River were similar to those from Da-han River. Also, the total concentrations of OCPs in Erh-jen River ranged from 0.08 to 8.20 $\mu\text{g/g}$ TOC with the mean concentration of 2.67 $\mu\text{g/g}$ TOC, which is similar to that in Da-han River. This result reflects that the contamination of OCPs in sediments from the selected rivers in Taiwan may be partly from the same pollution sources.

Fig. 3 illustrates the detection frequencies of OCPs in sediments collected from the selected rivers. Among the OCPs detected in Da-han River, endosulfan sulfate was the most frequent found compound in sediment (60%), followed by *p,p'*-DDE (55%) and dieldrin (55%). Also, high detection frequencies of *p,p'*-DDD and *p,p'*-DDT were demonstrated. This result shows that DDTs were the most dominant compounds in the sediments from Da-han River. Similar to the results in Da-han River, high detection frequencies of β -HCH, *p,p'*-DDD, endosulfan sulfate and *p,p'*-DDE were observed in sediments from Erh-jen River. The detection frequencies

Table 3
Concentrations of OCP residues in surface sediment of the Da-han River

Sample	OCPs (ng/g of dry sediment) ^a																
	α -BHC	β -BHC	Hept.	Hept. epoxide	Aldrin	Diel-drin	Endrin	Endo. I	Endo. II	Endo. sulfate	Endrin aldehyde	Endrin ketone	DDE	DDD	DDT	Methoxy.	Σ OCP
DH-101	1.96	<0.35	<0.15	<0.05	0.4	<0.22	<0.15	<0.08	<0.16	<0.13	<0.06	<0.13	1.33	0.69	0.75	<0.3	5.13
DH-201	3.74	<0.35	<0.15	0.08	<0.12	<0.22	<0.15	<0.08	<0.16	<0.13	0.32	<0.13	<0.12	<0.18	1.03	<0.3	5.17
DH-301	1.31	<0.35	<0.15	<0.05	<0.12	<0.22	<0.15	<0.08	0.18	<0.13	<0.06	<0.13	<0.12	<0.18	0.21	<0.3	1.7
DH-401	4.32	<0.35	<0.15	<0.05	0.29	<0.22	<0.15	2.41	1.36	2.21	<0.06	<0.13	1.5	<0.18	1.76	<0.3	13.85
DH-501	<0.12	<0.35	<0.15	<0.05	3.16	<0.22	<0.15	<0.08	<0.16	0.88	<0.06	<0.13	<0.12	<0.18	0.32	<0.3	4.36
DH-102	4.55	0.39	<0.15	0.15	5.8	<0.22	<0.15	<0.08	<0.16	<0.13	<0.06	<0.13	3.89	<0.18	<0.18	<0.3	14.78
DH-202	<0.12	<0.35	<0.15	<0.05	0.18	<0.22	<0.15	<0.08	0.03	<0.13	<0.06	<0.13	2.29	<0.18	0.83	<0.3	3.33
DH-302	<0.12	<0.35	<0.15	<0.05	<0.12	<0.22	<0.15	<0.08	<0.16	<0.13	<0.06	<0.13	<0.12	<0.18	0.47	<0.3	0.47
DH-402	<0.12	<0.35	<0.15	<0.05	0.89	<0.22	<0.15	<0.08	<0.16	<0.13	<0.06	<0.13	0.46	<0.18	<0.18	<0.3	1.35
DH-502	<0.12	<0.35	<0.15	<0.05	<0.12	<0.22	<0.15	<0.08	<0.16	<0.13	<0.06	<0.13	<0.12	<0.18	<0.18	0.2	0.2
DH-103	<0.12	<0.35	<0.15	<0.05	0.82	<0.22	<0.15	0.69	<0.16	0.8	<0.06	<0.13	0.51	0.56	<0.18	0.52	3.9
DH-203	0.73	<0.35	0.58	<0.05	<0.12	<0.22	<0.15	0.71	<0.16	0.62	<0.06	<0.13	1.52	<0.18	<0.18	>0.3	4.16
DH-303	<0.12	<0.35	<0.15	<0.05	<0.12	0.56	<0.15	<0.08	0.51	2.27	0.68	<0.13	<0.12	1.5	1.57	0.76	7.85
DH-403	<0.12	<0.35	1.15	<0.05	<0.12	<0.22	<0.15	<0.08	<0.16	3.18	2.39	<0.13	<0.12	1.36	2.26	0.72	11.06
DH-503	1.44	1.08	0.54	<0.05	<0.12	<0.22	<0.15	<0.08	<0.16	2.03	<0.06	<0.13	<0.12	3.34	<0.18	<0.3	8.43
DH-104	<0.12	<0.35	1.23	<0.05	0.8	<0.22	<0.15	1.08	0.62	0.89	0.66	0.57	0.72	1.07	<0.18	<0.3	7.64
DH-204	<0.12	<0.35	<0.15	<0.05	<0.12	<0.22	<0.15	1.54	0.85	2.16	<0.06	<0.13	0.63	0.91	<0.18	<0.3	6.09
DH-304	<0.12	0.73	0.63	<0.05	0.59	<0.22	<0.15	<0.08	<0.16	1.67	0.56	<0.13	<0.12	0.64	<0.18	<0.3	4.82
DH-404	1.9	<0.35	1.57	<0.05	0.72	<0.22	<0.15	<0.08	<0.16	2.66	1.07	<0.13	1.22	1.65	2.64	<0.3	13.43
DH-504	<0.12	0.67	<0.15	<0.05	0.52	<0.22	<0.15	0.57	<0.16	1.38	0.89	1.34	0.71	1.41	<0.18	2.85	10.34

^a Hept.: heptachlor; Endo.: endosulfan, Methoxy.: methoxychlor.

Table 4
Concentrations of OCP residues in surface sediment of the Erh-jen River

Sample	OCPs (ng/g of dry sediment) ^a																
	α -BHC	β -BHC	Hept.	Hept. epoxide	Aldrin	Dieldrin	Endrin	Endo. I	Endo. II	Endo. sulfate	Endrin aldehyde	Endrin ketone	DDE	DDD	DDT	Methoxy.	Σ OCP
EJ-101	<0.12	0.57	<0.15	<0.24	<0.05	<0.12	<0.22	<0.08	<0.16	<0.13	<0.06	<0.13	<0.12	<0.18	<0.18	<0.3	0.57
EJ-201	<0.12	<0.35	<0.15	<0.24	<0.05	<0.12	<0.22	<0.08	<0.16	<0.13	0.89	<0.13	<0.12	<0.18	2.35	<0.3	3.24
EJ-301	<0.12	0.53	<0.15	<0.24	<0.05	<0.2	<0.22	<0.08	<0.16	<0.13	<0.06	<0.13	<0.12	<0.18	0.09	0.57	1.19
EJ-401	<0.12	5.35	<0.15	<0.24	<0.05	<0.12	<0.22	<0.08	<0.16	1.27	<0.06	<0.13	<0.12	<0.18	<0.18	<0.3	6.62
EJ-501	<0.12	2.27	<0.15	<0.24	<0.05	1.29	<0.22	<0.08	<0.16	<0.13	<0.06	<0.13	0.38	<0.18	0.57	<0.3	4.51
EJ-102	0.97	5.93	<0.15	<0.24	<0.24	<0.12	0.64	<0.08	<0.16	<0.13	<0.06	<0.13	0.28	<0.18	1.44	<0.3	9.26
EJ-202	<0.12	2.61	<0.15	<0.24	<0.05	<0.12	0.23	0.63	<0.16	0.21	<0.06	<0.13	<0.12	0.25	0.3	<0.3	4.23
EJ-302	<0.12	5.09	<0.15	<0.24	<0.05	<0.12	<0.22	0.5	<0.16	<0.13	0.02	<0.13	<0.12	2.28	0.71	<0.3	8.60
EJ-402	<0.12	7.15	<0.15	<0.24	<0.05	0.17	<0.22	0.39	2.12	<0.13	<0.06	<0.13	0.4	0.45	0.29	<0.3	10.97
EJ-502	<0.12	14.1	5.61	<0.24	<0.05	<0.12	<0.22	1.71	0.71	<0.13	0.2	<0.13	<0.2	0.26	1.08	<0.3	23.67
EJ-103	1.17	3.77	0.64	<0.24	<0.05	<0.12	<0.22	2.12	<0.16	1.25	0.64	1.77	0.66	2.7	<0.18	0.74	15.46
EJ-203	<0.12	<0.35	<0.15	<0.24	<0.05	<0.12	<0.22	<0.08	<0.16	0.96	<0.06	<0.13	<0.12	0.75	<0.18	1.14	2.85
EJ-303	0.84	4.4	0.51	<0.24	<0.05	0.58	<0.22	2.3	<0.16	0.72	1.99	1.02	0.75	3.9	<0.18	1.23	18.24
EJ-403	1.11	3.1	0.95	<0.24	<0.05	0.53	<0.22	<0.08	<0.16	1.58	<0.06	<0.13	1.69	1.55	5.57	<0.3	16.08
EJ-503	1.5	<0.35	0.6	<0.24	<0.05	0.77	<0.22	3.26	3.01	0.89	1.42	<0.13	1.39	1.09	<0.18	<0.3	13.93
EJ-104	1.36	1.63	<0.15	<0.24	<0.05	1.09	<0.22	1.07	0.93	1.12	0.61	<0.13	0.65	1.14	<0.18	1.71	11.31
EJ-204	1.37	1.48	<0.15	<0.24	<0.05	0.98	<0.22	0.52	<0.16	1.52	<0.06	1.76	0.86	3.49	<0.18	7.39	19.37
EJ-304	1.2	5.02	<0.15	<0.24	<0.05	0.58	<0.22	1.13	<0.16	0.6	<0.06	<0.13	0.67	1.9	1	<0.3	12.1
EJ-404	1.29	2.83	0.65	<0.24	<0.05	<0.12	<0.22	<0.08	<0.16	1.72	<0.06	<0.13	1.1	1.66	1.04	0.71	11
EJ-504	1.18	<0.35	<0.15	<0.24	<0.05	<0.12	<0.22	1.09	<0.16	0.37	0.66	0.61	0.78	<0.18	<0.18	<0.3	4.69

^a Hept.: heptachlor; Endo.: endosulfan, Methoxy.: methoxychlor.

Table 5

Concentration ranges, medians and mean values of OCP residues in surface sediments collected from Da-han and Erh-jen Rivers

OCPs	Da-han river ($\mu\text{g/g TOC}$)			Erh-jen River ($\mu\text{g/g TOC}$)		
	Range	Median	Mean	Range	Median	Mean
Σ OCP	0.21–7.42	1.73	2.34	0.08–8.20	2.03	2.67
Σ HCH	ND–2.08	0.13	0.39	ND–3.47	0.70	1.03
Σ Cyclodiene	ND–4.51	0.87	1.05	ND–5.8	0.53	0.92
Σ DDT	0.06–2.91	0.43	0.90	ND–2.03	0.50	0.72

ND: not detected.

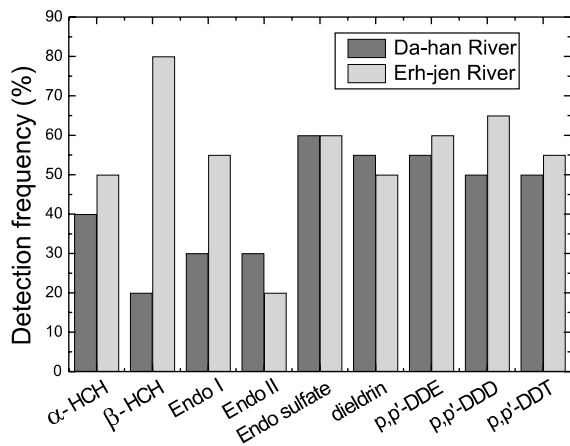


Fig. 3. The detection frequencies of OCPs in the sediments collected from Da-han and Erh-jen Rivers, Taiwan. The results are the means of the OCPs from the five sampling stations and four sampling times in each river.

were 80%, 65%, 60% and 60%, respectively. These results clearly show that DDTs and HCHs were the main OCPs used in Taiwan.

3.3. Characteristics of OCP contamination in sediments

The relative concentrations of the parent DDT compound and its metabolites can provide useful information on the pollution source. The compound *p,p'*-DDT is the active ingredient of the DDT pesticides and typically makes up approximately 80% of the technical formulation (Bopp et al., 1982; Hendy and Peake, 1996). Fig. 4 illustrates the ratios of DDD/DDE and $(\text{DDE} + \text{DDD})/\Sigma \text{DDT}$ in the sediments from Da-han and Erh-jen Rivers. Ratios of $(\text{DDE} + \text{DDD})/\Sigma \text{DDT}$ in sediments ranged from 0.37 to 1.0. Since ratio of $(\text{DDE} + \text{DDD})/\Sigma \text{DDT} > 0.5$ is reported to be subjected to long-term weathering (Hites and Day, 1992), results obtained in this study indicate that the DDT compounds in sediments may be mainly derived from DDT-treated aged and weathered agricultural soils. Also, the ratios of DDD/DDE ranged from 0.24 to 1.99 and from 0.78 to

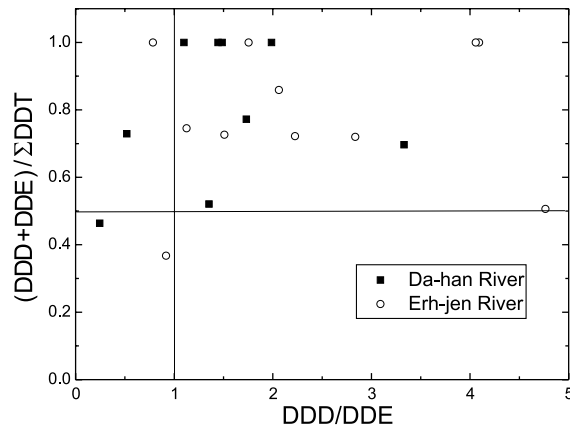


Fig. 4. The relationship between DDD/DDE and $(\text{DDE} + \text{DDD})/\Sigma \text{DDT}$ in sediments from Da-han and Erh-jen Rivers.

5.20 for Da-han and Erh-jen Rivers, respectively. Most of the ratios are greater than unity. This means that the total DDT found in sediments from the selected rivers is dominated by *p,p'*-DDD, the product of anaerobic degradation of *p,p'*-DDT. This result is consistent with the measured ORP values (−372 to 184 mV), suggesting that the input of DDT compounds to the rivers was via the weathered agricultural soils and was retained under anaerobic conditions within the sediment in Taiwan rivers.

The frequency of detection of HCH compounds in environmental samples showed that the contamination of HCH was also widespread in Taiwan rivers. Recent studies also show that the contamination of HCH isomers is a serious problem worldwide (Fellin et al., 1996; Sarkar et al., 1997; Hong et al., 1999; Walker et al., 1999). HCH pesticide is an inexpensive insecticide and had been used for agricultural purpose to control the insects in fruit, grain and vegetable crops in Taiwan since 1950s. Technical-grade HCH consists principally of five isomers, α -HCH (60–70%), β -HCH (5–12%), γ -HCH (10–15%), δ -HCH (6–10%) and ϵ -HCH (3–4%) (Walker et al., 1999). The most often found isomers in the environment are α -, β - and γ -isomers. α -HCH was found to be the dominant compound of HCH in the

sediments from Da-han River. Since the catchment of Da-han River is a heavily urbanized area, the contamination of α -HCH may be attributed to the long distance transport from other areas. Some OCPs such as endosulfan, heptachlor and HCH are still used in some developing countries around the tropical belt and may be transported through the atmosphere and gradually deposit in the river at higher latitudes. Oehme et al. (1996) measured the seasonal concentration changes of organochlorines in the European Arctic and found that long-range atmospheric transport from more polluted areas might lead to a significant concentration change in the Arctic air. It is estimated that the total global usage of technical grade HCH between 1948 and 1997 was around 10 million tons (Li, 1999). Recently, considerable unused stockpiles of technical-grade HCH was found in dump sites in Africa and Near East (FAO, 1998). Some of the containers have damaged and are leaking. Among the HCH isomers, α -HCH is more likely to partition to the air and transport for a long distance. A study conducted in India examining the flux of HCH also indicated that most of the HCH applied annually was lost to the atmosphere (Takeoka et al., 1991). These results depict that α -HCH may be mainly from the long-range transport (LRT). Since α -HCH exhibits the most carcinogenic activity among HCH isomers, the contamination levels detected may pose a high ecotoxicity for aquatic organisms.

Different HCH contamination patterns for Erh-jen River were observed. The detection frequencies of α - and β -HCH were 50% and 80%, respectively. The similar detection frequency of α -HCH in the selected rivers shows that the LRT may be the major source of α -HCH in Erh-jen River. Unlike the α -HCH, β -HCH is more lipophilic and is the predominant isomer in soils and animal tissue and fluids (Willett et al., 1998). Also, β -HCH is resistant to hydrolysis and environmental degradation. Since Erh-jen River is a strong agricultural sector in southern Taiwan, the abundance of the β -HCH may be attributed to the weathered soils, which is also consistent with the results of Σ DDT.

4. Conclusions

The results obtained in this study documents the first known analysis of OCP distributions in sediments collected from Da-han and Erh-jen Rivers, Taiwan. There still exist a variety of OCP residues in the river sediments in Taiwan. The total concentrations of OCPs in sediments ranged from 0.2 to 14.8 and 0.6 to 29.5 ng/g dry weight for Da-han and Erh-jen Rivers, respectively, which corresponds to 0.21–7.42 and 0.08–8.20 μ g/g TOC when normalized to TOC concentrations. DDTs and HCHs were the dominant OCPs in sediments from the

selected rivers. Da-han River was mainly contaminated with α -HCH and *p,p'*-DDE, while the main pesticide residues in Erh-jen River were β -HCH and *p,p'*-DDD. The detection frequencies of the metabolites were higher than those of parent compounds. This distribution pattern reflects the fact that the contamination of the collected sediments is mainly from LRT and weathered agricultural soils.

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